



Efficient Decolorization of Water and Oil-Soluble Azo Dyes by *Enterococcus avium* Treated with HP- β -CD

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ABSTRACT

The industrial effluents containing azo dyes will contaminate the environment without decolorization. Bacterial decolorization generally demonstrates good color removal effects. In this study, the capability of *Enterococcus avium* to decolorize both water and oil-soluble azo dyes were investigated. The bacteria was incubated under static conditions in the presence of 50mg/L Amaranth or 5 mg/L Sudan I in LB media at 37°C for 10 h, and reduction of the dyes was monitored. The preferable pH for the decolorization of Amaranth by ts17 strain was between pH 6.0-9.0, and with the optimum pH at 7.0. The promotion effect of NaCl on the decoloration of Amaranth up to 5% (w/w) content, and the inhibition action appeared at 7% NaCl. The maximum decolorization concentration of the Amaranth reached 1450mg/L. Mg²⁺ and Fe³⁺ could accelerate Amaranth decolorization, and Cu²⁺, Zn²⁺, Co²⁺ showed inhibition of Amaranth decolorization. Compared with the traditional method dissolving oil-soluble dye in DMSO, HP- β -CD was used to monitor the decolorization rate of the sudan I and appropriate molar ratio of HP- β -CD to dye (n_c:n_s=9:1) effectively improved the decolorization rate from 85.45% to 99.45% after 8 reaction.

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Authors' Contribution

WF designed the study and purchased materials. PS and HS performed the experimental work, analyzed the data and wrote the article. JS, RL and WX helped perform the analysis with constructive discussion.

Key words

Azo dyes, Decolorization, *Enterococcus avium*, HP- β -CD.

INTRODUCTION

Azo dyes represent a major group of dyes causing environmental concern because of their color, bio-recalcitrance, and potential toxicity to animal and human (Pierce *et al.*, 2003; Forgacs *et al.*, 2004). They reduce the amount of sunlight to photosynthetic organisms resulting in decreased oxygen levels in aquatic ecosystems (Yogesh *et al.*, 2013). The presence of dyes or their degradation products in water even at very low concentrations can be toxic and sometimes carcinogenic, mutagenic, or teratogenic to various organisms, including human being (Novotny *et al.*, 2006; Hai *et al.*, 2007). Development of an efficient dye degradation technology requires application of suitable strains and its use under favorable conditions to realize its degradation potential (Novotny and Nenov, 2004). Most studies on azo dye biodegradation have focused on bacteria and fungi, where bacteria are found to be more efficient (Elisangela *et al.*, 2009; Vijaykumar *et al.*, 2007; Aftab *et al.*, 2011).

Cyclodextrins (CDs) are cyclic (α -1, 4)-linked oligosaccharides with different numbers of α -D-glucopyranose units (Larsen *et al.*, 1998). Cyclodextrins have a toroidal or cone shape due to the lack of freely rotating bonds between glucopyranose units and possess a hydrophilic outer surface and a hydrophobic inner cavity in which hydrophobic small guest molecules can reside. In addition to this hydrophobic interaction, other interactions such as Van der Waals forces, electrostatic interactions, and hydrogen bonding interactions can also be driving forces for complex formation of guest molecules with CDs (Szejtli, 1998). This property provides CDs with the capacity to increase the aqueous solubility of hydrophobic drugs (Loftsson and Brewster, 1996; Stella and Rajewski, 1997). But it is rarely used to improve the solubility of oil soluble dyes. Amongst cyclodextrins, hydroxypropyl- β -cyclodextrin (HP- β -CD) possesses greater ability to accommodate several lipophilic molecules (Miro *et al.*, 2006).

In the present work, the dye decolorization efficiency of strain ts17 isolated from activated sludge is studied. Amaranth and sudan (Table I) were chosen to study the influence of different conditions on the decolorization ability of the bacteria and its ability of water and oil-soluble dye decoloring. Previously, it was demonstrated that

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water-soluble azo dyes can freely move in medium, pass through cell walls more easily than water-insoluble azo dyes, and are degraded by azoreductases in several other bacteria (Chen, 2006). Although a few reports described the ability of some bacteria to decolor both water and oil-soluble azo dyes (Chen *et al.*, 2009). Furthermore, no report regarding either the decolorization of both azo dyes by *Enterococcus avium* or its application to improve the decoloration rate of oil soluble dyes with HP- β -CD has been published.

Table I.- Chemical structures of water-soluble azo dye Amaranth and oil-soluble azo dye Sudan I used in this study.

Amaranth	Sudan I

MATERIALS AND METHODS

Chemicals and media

The dyes such as Amaranth and Sudan I used in this study were supplied by Aladdin chemical, China. The other chemicals used were of analytical grade.

The basal culture medium used for the dye decolorization studies were LB medium with azo dyes. Regarding the influence of culture conditions on dye decolorization, some basic conditions were changed and others remained unchanged.

Isolation and cultivation of microbes

The activated sludge, which was obtained from a plant sewage treatment located at Shandong Province, eastern China, was used as a source for the isolation of bacterial strains. The colony showing maximum decolorization zone named ts17 was transferred to the broth containing the same dye (50 mgL⁻¹) to test the decolorization potential of the bacterium in static liquid medium at 37°C and drawn growth curve of bacterium ts17 on the same condition.

16S rDNA sequencing

Extraction of genomic DNA was performed according to the method of Sambrook and Russell. PCR amplification of the 16S rRNA gene was performed using the following two

primers: 27F (5'-AGAGTTTGATCMTGGCTCAG-3') and 1492R (5'-GGTACCTTGTTACGACTT-3').

The purified gene fragment was sequenced by Shanghai Sangon Biological Engineering Technology and Services Co., Ltd. Phylogenetic neighbours were identified using GenBank (<http://www.ncbi.nlm.nih.gov/blast/>), and multiple alignments were generated using CLUSTAL X. A phylogenetic tree was constructed from the evolutionary distance data with the neighbour-joining, maximum-likelihood PHYLIP (phylogeny inference package, version 3.5c. Department of Genetics, University of Washington, Seattle) and maximum-parsimony methods using MEGA 5.0, and bootstrap values were calculated based on 1,000 replicates.

Decolorization of amaranth by growing cells

In order to study the effect of initial dye concentration on dye decolorization potential and tolerance of the strain ts17, the isolate was cultivated for more than 4 days and was amended with different concentrations (50-200mg/L) of Amaranth at concentration gradient 50mg/L. To study the effect of initial pH on dye decolorization, the strain ts17 was cultivated for 24 h at varying initial pH (5-10) and was supplemented with 50 mg/L of Amaranth. The effect of salt concentration on dye decolorization was studied by using different concentrations of sodium chloride (0%, 1%, 3%, 5% and 7%) in the medium with 50 mg/L of Amaranth. To inspect the influence of various metal ions such as Mg²⁺ (MgCl₂), Co²⁺ (CoCl₂), Zn²⁺ (ZnSO₄), Mn²⁺ (MnCl₂), Fe³⁺ (FeCl₃) and Cu²⁺ (CuSO₄) on decolorization activity, experiments were carried out in presence of these metals at concentration of 2 mM. Metal ions from stock solutions were added to the cell suspensions and incubated for 15 min, and then flasks were added with the dyes.

In the process of detecting decolorization rate affected by various kinds of factors, supernatants were analysed at 520 nm on spectrophotometer every 2 h. The decolorization rate (DR) was calculated by using the formula (Shekhar *et al.*, 2012):

$$\text{DR (\%)} = \frac{\text{Initial absorbance} - \text{Observed absorbance}}{\text{Initial absorbance}} \times 100 \dots (1)$$

Decolorization of Sudan I by strain ts17

In order to study the decolorizing ability and decoloring conditions of oil-soluble azo dyes by strain ts17, the bacteria were incubated in the presence of 5 mg/L Sudan I which was dissolved in DMSO broth at 37°C without agitation, and reduction of the dyes was monitored.

The time-course of Sudan I decolorization by ts17 was monitored using high-performance liquid chromatography (HPLC). The peak area was used to

calculate the concentration of Sudan I. Reduction of Sudan I was determined by monitoring the disappearance of the absorption peak for the dye at 500 nm (Chen *et al.*, 2009). The average of the same measurements group was used as initial or monitored peak area. The decolorization rate was calculated by using the following formula:

$$\text{DR (\%)} = \frac{\text{Initial peak area} - \text{Monitored peak area}}{\text{Initial peak area}} \times 100 \dots \dots (2)$$

In addition, we also explored the decolorization of Sudan I by the presence of different molar ratio of HP- β -CD. We mixed different molar ratio of HP- β -CD to Sudan I at 3:1, 6:1, 9:1, 12:1 to medium before inoculation, respectively.

RESULTS AND DISCUSSION

Decolorizing bacteria

Comparison of the near full-length 16S rRNA gene sequence (1,512 bp) from strain 17 with other bacteria showed that it is most closely affiliated to the genus *Enterococcus* (Fig. 1). The highest 16S rRNA gene

sequence similarity of 99 % was found with *Enterococcus avium* (GenBank accession number KP735772).

Decolourization of amarnath

When dye decolorization was conducted at shaking condition (120 rpm), no decolorization ability was found for strain ts17. This result demonstrated that the static condition was necessary for the decolorization of ts17, as it is also the case for most of the microbial species for the decolorization of azo dyes. The general mechanism for the azo dye degradation is through reduction of azo bond by azoreductase (Chacko and Subramaniam, 2011). It has been suggested that the azo bond cleavage is catalyzed by azoreductase, which is very sensitive to oxygen in the medium (Nikolova and Nenov, 2004). After decolourization of Amaranth by the bacterium, there was no visible dye adsorbed to the biomass and no colouration in the methanol extract of the biomass was found. This indicates that the observed dye decolourization is mainly due to the biological activity of the bacterium rather than adsorption or biosorption.

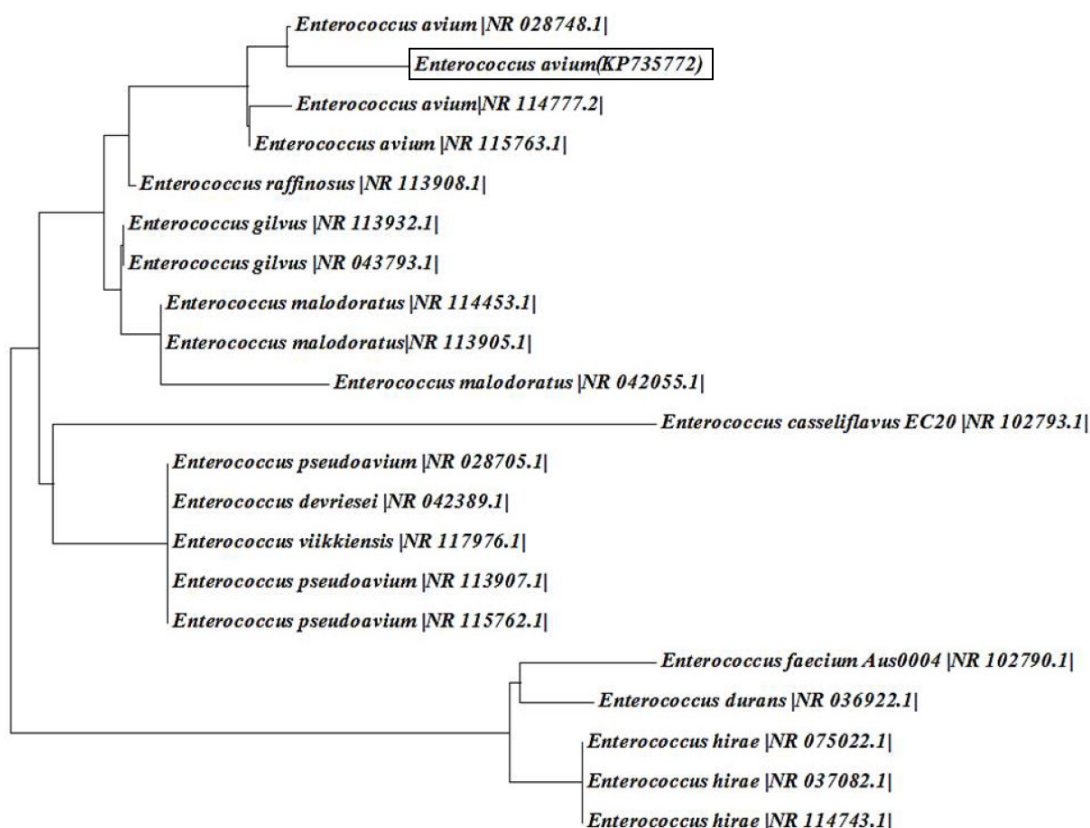


Fig. 1. Neighbour-joining phylogenetic tree based on 16S rRNA gene sequences showing the positions of strain ts1-17 (GenBank accession number KP735772) and related taxa. Bar=0.001 substitutions per nucleotide position.

Ability of ts17 for decoloring Amaranth was checked at higher dye concentrations. The dye was also decolorized when dye concentration reached 1450 mg/L. Longer time is required for decolorization when dye concentration is increased.

Bacterial growth curve of ts17 is shown in Figure 2, which indicates 0-4 h of lag phase, 4-10 h of logarithmic phase and 12 h of stationary phase.

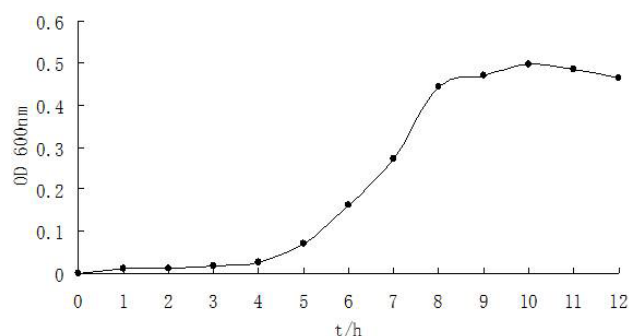


Fig. 2. The cell growth curve of bacterium ts17 with 50mg/L Amaranth in 12 h.

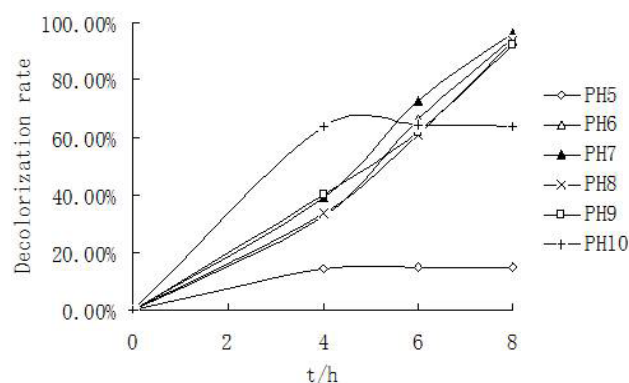


Fig. 3. Effects of pH on decolorization rate of Amaranth by strain ts17 in 8 h.

Effect of pH

As shown in Figure 3, the decolorization rate of the incubation are 14.99%, 94.57%, 96.21%, 93.67%, 91.95% and 63.93% respectively, corresponding to pH at 5.0, 6.0, 7.0, 8.0, 9.0 and 10.0 after 8 h incubation. The preferable pH for the decolorization of Amaranth azo dye by ts17 strain was between pH 6.0-9.0, and with optimum pH of 7.0.

Effect of NaCl concentration

The concentrations of NaCl have impacts on the decolorization of Amaranth by strain ts17. After 6 h of incubation, the decolorization rate of the incubation

are 54.07%, 84.10 %, 96.82%, 72.26% and 32.35%, respectively, corresponding to 0%, 1%, 3%, 5% and 7% NaCl solutions (Fig. 4), indicating the promotion effect of NaCl on the decolorization up to 5% (w/w) content, and the inhibition action appeared at 7% NaCl.

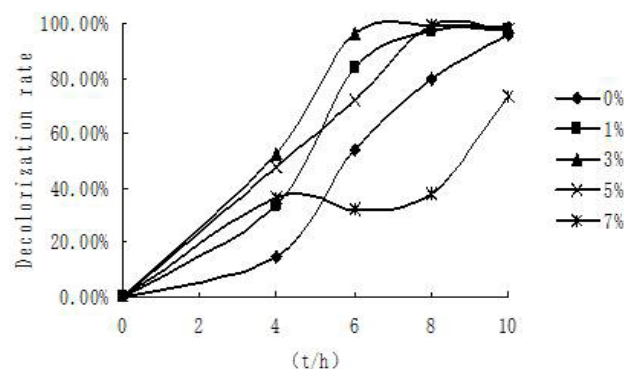


Fig. 4. Effects of NaCl% on decolorization rate of Amaranth by strain ts17 in 10 h.

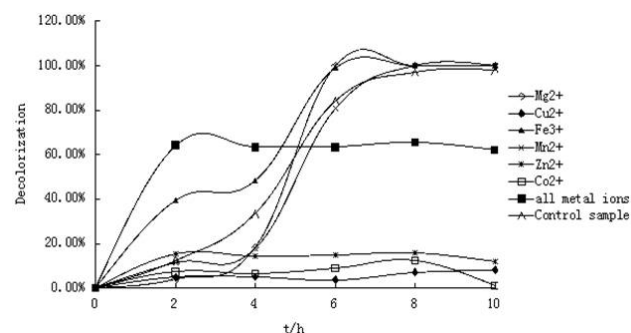


Fig. 5. Effects of heavy metal ions on decolorization rate of Amaranth by strain ts17 in 10 h.

Effect of heavy metal ions

As shown in Figure 5, the decolorization rate of Amaranth was 100% and 99.08% in the presence of Mg^{2+} and Fe^{3+} , respectively, compared with that of the control sample of 84.10% after 6 h. Fe^{3+} can obviously improve the decolorization rate in the lag phase of bacteria reflected in bacterial growth curve. Moreover, we detected that decolorization rate increased from 18.36% to 100% between 4 h and 6 h with Mg^{2+} . Mg^{2+} can obviously improve the decolorization in the logarithmic phase of bacteria, and the contributions of decolorization rate account for 80%. Mg^{2+} has been reported to enhance the enzyme activity of azoreductase from *Rhodobacter sphaeroides* (Yan *et al.*, 2004). On the other hand, other metal ions including Cu^{2+} , Zn^{2+} and Co^{2+} have inhibitory effect on the decolorization

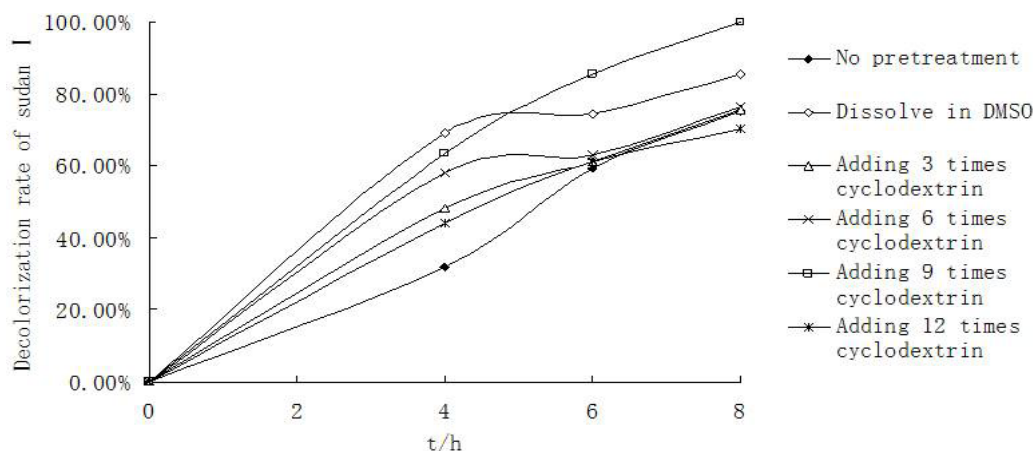


Fig. 6. High-performance liquid chromatography (HPLC) analysis of Sudan I decolorization after different pretreatments by strain ts17 in 8 h.

process, and the decolorization rate was only 7.10%, 15.89% and 12.46%, respectively after 8 h of process at a concentration of 2 mM, which is much lower than that of the control samples of 97.04%. Shekhar *et al.* (2012) reported a strong negative effect of Zn^{2+} and Co^{2+} at 5 mM concentration. In another study, less than 30 % residual enzyme activity for Zn^{2+} was found in the range from 5 to 50 mM (Yang *et al.*, 2013). Mn^{2+} did not show any major effect on decolorization of Amaranth. When all metal ions were added, decolorization rate reached 60% in 2 h and remained unchanged afterwards.

Decolourization of Sudan I dyes

It was found that ts17 also has ability of decolorizing oil solution azo dye such as Sudan I and decolorization rates of 69.09%, 74.55%, and 85.45% at 4 h, 6 h and 8 h when dissolved in DMSO, respectively (Fig. 6). These data were higher than that of control which showed decolorization rate of 32.12%, 59.39%, 75.15%, respectively at the same time period.

In the cell suspensions of ts17, the Sudan I dye and HP- β -CD are simultaneously added to the solution in the range of 3-9 mole ratios (Fig. 6). The addition of the HP- β -CD had a positive effect on the decolorization rate of ts17 strain on the Sudan I dye, and with the increasing concentration of HP- β -CD, decolorization rates of Sudan I increased. However, further increase in the molar ratio of HP- β -CD in Sudan I decreased the decolorization rate. So, the treatment of Sudan I with 9 times mole ratio of HP- β -CD increased the decolorization rate comparable to the dye dissolved in DMSO.

So, in brief the isolated strain ts17 has a high ability to decolorize azo dye Amaranth. The ts17 cultured in LB medium under static condition containing Amaranth (50

mg/L) decolored the dyes more than 99% at 37°C, pH7.0, with 3% NaCl after 8 h. The maximum decolorization concentration of the dye reached 1450mg/L. Mg^{2+} and Fe^{3+} accelerated dye decolorization and Cu^{2+} , Zn^{2+} , Co^{2+} on the other hand inhibited dye decolorization. The strain ts17 was also able to completely reduce the oil-soluble dyes Sudan I. It was also shown that the appropriate concentration of HP- β -CD mixed with dye ($n_c:n_s=9:1$) effectively improved the decolorization rate from 85.45% to 99.45% after 8h.

CONCLUSION

In conclusion, *Enterococcus avium* is capable of efficiently decoloring monoazo water-soluble Amaranth and monoazo oil-soluble Sudan I treated with HP- β -CD. Our further investigations will be focused on the mechanism by which the HP- β -CD played and the degradation of aromatic amines caused by azo dye reduction.

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Statement of conflict of interest

The authors declare that there is no conflict of interests regarding the publication of this article

REFERENCES

- Aftab, U., Khan, M.R., Mahfooz, M., Ali, M., Aslam, S.H. and Rehman, A, 2011. Decolourization and

- degradation of textile azo dyes by corynebacterium sp. isolated from industrial effluent. *Pakistan J. Zool.*, **43**: 1-8.
- Chacko, J. and Subramaniam, K., 2011. Enzymatic degradation of azo dyes - A review. *Int. J. environ. Sci.*, **1**: 1250-1260.
- Chen, H., 2006. Recent advances in azo dye degrading enzyme research. *Curr. Protein Pept. Sci.*, **7**: 101-111. <https://doi.org/10.2174/138920306776359786>
- Chen, H., Xu, H., Heinze, T.M. and Cerniglia, C.E., 2009. Decolorization of water and oil-soluble azo dyes by *Lactobacillus acidophilus* and *Lactobacillus fermentum*. *J. Ind. Microbiol. Biotechnol.*, **36**: 1459-1466. <https://doi.org/10.1007/s10295-009-0633-9>
- Elisangela, F., Andrea, Z., Fabio, D.G., Cristiano R.M., Regina D.L. and Artur, C.P., 2009. Biodegradation of textile azo dyes by a facultative *Staphylococcus arlettae* strain VN-11 using a sequential microaerophilic/aerobic process. *Int. Biodeterior. Biodegrad.*, **63**: 280-288. <https://doi.org/10.1016/j.ibiod.2008.10.003>
- Forgacs, E., Cserhati, T. and Oros, G., 2004. Removal of synthetic dyes from wastewaters: A review. *Environ. Int.*, **30**: 953-971. <https://doi.org/10.1016/j.envint.2004.02.001>
- Hai, F., Yamamoto, K. and Fukushi, K., 2007. Hybrid treatment system for dye wastewater. *Crit. Rev. environ. Sci. Technol.*, **37**: 315-377. <https://doi.org/10.1080/10643380601174723>
- Larsen, K.L., Duedahl-Olesen, L., Christensen, S.J.H., Mathiesen, F., Pedersen, L. and Zimmermann, W., 1998. Purification and characterization of a cyclodextrin glycosyltransferase from *Paenibacillus* sp. *Carbohydr. Res.*, **310**: 211-219. [https://doi.org/10.1016/S0008-6215\(98\)00178-5](https://doi.org/10.1016/S0008-6215(98)00178-5)
- Lofstsson, T. and Brewster, M.E., 1996. Pharmaceutical applications of cyclodextrins. 1. Drug solubilization and stabilization. *J. Pharm. Sci.*, **85**: 1017-1025. <https://doi.org/10.1021/js950534b>
- Miro, A., Quaglia, F., Giannini, L., Cappello, B. and Giannini, L., 2006. Drug/cyclodextrin solid systems in the design of hydrophilic matrices: A strategy to modulate drug delivery rate. *Curr. Drug Deliv.*, **3**: 373-378. <https://doi.org/10.2174/156720106778558994>
- Nikolova, N. and Nenov, V., 2004. Azo dye Schwarz GRS bioconversion under various conditions. *Water Air Soil Pollut.*, **4**: 137-146. <https://doi.org/10.1023/B:WAFO.0000044793.39140.c0>
- Novotny, C., Dias, N., Kapanen, A., Malachova, K., Vandrovova, M., Itavaara, M. and Lima, N., 2006. Comparative use of bacterial, algal and protozoan tests to study toxicity of azo and anthraquinone dyes. *Chemosphere*, **63**: 1436-1442. <https://doi.org/10.1016/j.chemosphere.2005.10.002>
- Novotny, C., Svobodova, K., Kasinath, A. and Erbanova, P., 2004. Biodegradation of synthetic dyes by *Irpex lacteus* under various growth conditions. *Int. Biodeter. Biodegr.*, **54**: 215-223. <https://doi.org/10.1016/j.ibiod.2004.06.003>
- Pierce, C.I., Lloyd, J.R. and Bhattacharya, G.J.T., 2003. The removal of color from textile wastewater using whole bacterial cells: A review. *Dyes Pigments*, **58**: 179-185. [https://doi.org/10.1016/S0143-7208\(03\)00064-0](https://doi.org/10.1016/S0143-7208(03)00064-0)
- Shekhar, B.J., Shripad, N.S., Dayanand, C.K., Ranjit, G.G. and Jyoti, P.J., 2012. Biodecolorization of azo dye remazol orange by *Pseudomonas aeruginosa* BCH and toxicity (oxidative stress) reduction in *Allium cepa* root cells. *Appl. Biochem. Biotechnol.*, **168**: 1319-1334. <https://doi.org/10.1007/s12010-012-9860-z>
- Stella, V.J. and Rajewski, R.A., 1997. Cyclodextrins: their future in drug formulation and delivery. *Pharmaceut. Res.*, **14**: 556-567. <https://doi.org/10.1023/A:1012136608249>
- Szejtli, J., 1998. Introduction and general overview of cyclodextrin chemistry. *Chem. Rev.*, **98**: 1743-1753. <https://doi.org/10.1021/cr970022c>
- Vijaykumar, M.H., Vaishampayan, P.A., Shouche, Y.S. and Karegoudar, T.B., 2007. Decolourization of naphthalene-containing sulfonated azo dyes by *Kerstersia* sp. strain VKY1. *Enzyme Microb. Technol.*, **40**: 204-211. <https://doi.org/10.1016/j.enzmictec.2006.04.001>
- Yan, B., Zhou, J., Wang, J., Du, C., Hou, H., Song, Z. and Bao, Y., 2004. Expression and characteristics of the gene encoding azoreductase from *Rhodobacter sphaeroides* AS1.1737. *FEMS Microbiol. Lett.*, **236**: 129-136. <https://doi.org/10.1111/j.1574-6968.2004.tb09638.x>
- Yang, Y., Lu, L., Gao, F. and Zhao, Y., 2013. Characterization of an efficient catalytic and organic solvent-tolerant azoreductase toward methyl red from *Shewanella oneidensis* MR-1. *Environ. Sci. Pollut. Res.*, **20**: 3232-3239. <https://doi.org/10.1007/s11356-012-1221-5>
- Yogesh, M.K., Pallavi, D.K. and Vijay, L.M., 2013. Effective bioremoval and detoxification of textile dye mixture by *Alishewanella* sp. KMK6. *Appl. Microbiol. Biotechnol.*, **97**: 881-889. <https://doi.org/10.1007/s00253-012-3983-6>